



Tamires de Oliveira
Andrade

**AN INTEGRATIVE APPROACH TO DELIMIT SPECIES IN THE
EULAEMA CINGULATA AND *EULAEMA PSEUDOCINGULATA*
PAIR (HYMENOPTERA: APIDAE) USING MORPHOMETRIC AND
MOLECULAR EVIDENCE**

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APIDAE) UTILIZANDO EVIDÊNCIA MORFOMÉTRICA E MOLECULAR

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molecular**

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Dissertation submitted to the Graduate Program of the Museu de Zoologia da Universidade de São Paulo in partial fulfillment of the requirements for the degree of Master of Science (Systematics, Animal Taxonomy and Biodiversity).

Advisor: Prof. Dr. Carlos Roberto Ferreira Brandão

Co-advisor: Dr. Kelli dos Santos Ramos

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To my family for all support and trust. All thanks and this dedicatory will still be insufficient.

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“Penso no que faço, com fé. Faço o que tenho que fazer, com amor. Eu me esforço para ser melhor, pois bondade também se aprende. Mesmo quando tudo parecer desabar cabe a mim decidir entre rir e chorar, ir ou ficar, desistir ou lutar; porque descobri, no caminho incerto da vida que o mais importante é decidir.” Cora Coralina

ABSTRACT

The study of species delimitation has been a rich scientific field that makes use of different sources of data in order to identify independently evolving lineages that might be recognized as species. Males of the solitary orchid bees *Eulaema cingulata* (Fabricius, 1804) and *Eulaema pseudocingulata* Oliveira, 2006 are morphologically similar, although differ in the shape of the velvety area of the middle leg. These nominal species are relatively commonly recorded in bee diversity studies in Brazil; they have been, however, recognized by conflicting hypotheses of interspecific delimitation. Here we investigate the limits of both nominal species combining distinct methods: geometric morphometrics (GM), phylogenetics, COI-barcoding and pairwise genetic p-distance with a species delimitation method based on multispecies coalescent. We obtained data from 126 representative specimens of the entire geographical range in which the nominal species occur and assess species boundaries under the general lineage concept. We found substantial overlapping in the shape of forewings and heads of species in the GM analysis. The results of COI-barcoding analysis showed pair-wise genetic distances lower than 3% (within *E. cingulata* 1.3% and within *E. pseudocingulata* 0.7%, while between the two species 0.95%). The Bayesian phylogenetic reconstruction clustered both species in a strongly supported monophyletic group, however, these forms were not reconstructed as distinct clades. The sequencing of UCEs recovered 2.180 homologous loci that provided no sufficient variability to recognize both forms either as different species or as independent evolutionary units. Our results support the recognition of *E. cingulata* and *E. pseudocingulata* as members of the same evolutionary unit.

Keywords: Barcoding. Euglossini. Taxonomy. Orchid Bees. Phylogenomics. Species Delimitation. Ultraconserved Elements.

RESUMO

O estudo de delimitação de espécies tem sido um fértil campo da ciência, que faz uso de diferentes fontes de dados para identificar linhagens evolutivas independentes que devam ser reconhecidas como espécies. Os machos das espécies de abelhas solitárias das orquídeas *Eulaema cingulata* (Fabricius, 1804) e *Eulaema pseudocingulata* Oliveira, 2006 têm sido reconhecidas como entidades diferentes na literatura, são morfologicamente similares e se diferenciariam pela forma da região aveludada da perna média. Registros dessas espécies nominais são relativamente comuns em estudos de diversidade de abelhas no Brasil; no entanto, estes nomes têm sido reconhecidos a partir de hipóteses conflitantes de delimitação interespecífica. Aqui nós investigamos se existem limites entre ambas as espécies combinando métodos como Morfometria Geométrica (MG), filogenética, COI-barcoding e distância genética com um método de delimitação de espécies baseado em coalescência de múltiplas espécies. Obtivemos dados de 126 espécimes representativos de toda a distribuição geográfica em que as espécies nominais ocorrem e avaliamos os limites das espécies sob o conceito de linhagem geral. Encontramos sobreposições substanciais na forma das asas anteriores e das cabeças na análise de MG. Os resultados da análise do COI-barcoding mostraram distâncias genéticas inferiores a 3% (dentro de *E. cingulata* 1,3% e dentro de *E. pseudocingulata* 0,7%, enquanto entre as duas espécies 0,95%). A reconstrução filogenética bayesiana agrupou ambas as formas em um grupo monofilético fortemente suportado, apesar delas não terem sido reconstruídas, entretanto, como clados distintos. O sequenciamento de UCEs recuperou 2.180 loci homólogos que corroboram com análises anteriores e não forneceram variabilidade suficiente para reconhecer ambas as formas como espécies biológicas diferentes ou como unidades evolutivas independentes. Nossos resultados apoiam o reconhecimento de ambas as formas como membros de uma mesma unidade evolutiva.

Palavras-chave: Abelhas das Orquídeas. Barcoding. Delimitação de Espécies. Elementos Ultraconservados. Euglossini. Filogenômica. Taxonomia.

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1. INTRODUCTION

Taxonomic identification is a critical step in biodiversity studies and should be the starting point for further detailed research, since species' misidentifications or the application of ambiguous species concepts can carry serious negative impacts, hindering the investigation of any biological phenomenon (Austen *et al.*, 2016). Species that diverged recently may exhibit a very similar morphology and can be misclassified under an erroneous species name, leading to an underestimation of biological diversity (Leliaert *et al.*, 2014). On the other hand, conspecific individuals can display discrete variation in morphological characters (e. g. Quezada *et al.*, 2015; Lepeco & Gonçalves, 2018). When these variable characters are used for species identification, the same evolutionary unit may be assigned to different species, overestimating the biodiversity (Sivogini *et al.*, 2016).

To circumvent some of the limitations of taxonomy by using exclusively morphology (Padial *et al.*, 2010; Jorger & Schorodi, 2013), integrative taxonomy incorporates other approaches, such as geometric morphometrics, to assess intra- and inter-specific variation to facilitate the recognition of different species (Francisco *et al.*, 2008; Francoy *et al.*, 2012; Quezada-Euán *et al.*, 2015), and molecular biology techniques to complement species identification and delimitation based before solely on morphology (Blaxter, 2004; Goldstein & DeSalle, 2010). Molecular markers are an alternative tool for the characterization of different species (Blaxter, 2004). The identification of animal biological diversity using molecular markers has been mainly demonstrated through the use of the mitochondrial gene cytochrome c oxidase subunit I (COI). This gene has been used as a genetic barcode to identify taxa because it can be easily amplified by using universal primers (Herbert *et al.*, 2003), displays relatively high substitution rate suitable for constructing a phylogenetic tree (Galtier *et al.*, 2009), suffers comparatively rapid evolution of differentiation between closely related species, and is likely involved in the process of speciation (Hill, 2016). DNA barcoding is a common molecular method in the identification and delimitation of species from various groups, from invertebrates (Freitas *et al.*, 2018; Grando *et al.*, 2018) to birds (Herbert *et al.*, 2004). This method gained increased acceptance due to being simple and affordable, and to promise a practical, standardized, species-level identification tool that can be used for biodiversity assessment (Kress *et al.*, 2005; Padial & De La Riva, 2007). In a species identification analysis using a barcoding approach, within-species DNA sequences need to be more similar to one another than to sequences in different species (Ward *et al.*, 2005). Genetic distances between COI

sequences from individuals are calculated and values higher than 3% divergence has been commonly used as indicative of the presence of different species (Herbert *et al.*, 2003).

The application of DNA barcoding approach to systematics has revolutionized the discovery and description of biodiversity and its incorporation to taxonomic classification can help to solve some of the problems of relying only on morphological classification (Grando *et al.*, 2018). Basing a species description on a variety of characters from different and independent datasets is generally regarded as the best practice (Dayrat, 2005; DeSalle *et al.*, 2005). Integrative taxonomy is capable to recognize species which are identifiable by just a few apparent morphological criteria once their existence is revealed by other methods (Lajus *et al.*, 2015; Sáez & Lozano, 2005). When species are considered as independently evolving lineages (de Queiroz, 2007), different lines of evidence are additive to each other. Molecular evidence can provide an independent test of morphological assessments of species identity and vice versa (Page *et al.*, 2005). These combined approaches are a powerful tool to detect cryptic species, to clarify species boundaries, to estimate species diversity, including of bees (e.g. Gibbs, 2009; González-Vaquero *et al.*, 2019; Pauly *et al.*, 2019), and can also contribute to the estimation of true richness and hence to synonymize illegitimate nominal species (e. g. González-Vaquero & Roig-Alsina, 2019).

Evolutionary inference from single-locus data presents some limitations, however, including the retention of ancestral polymorphism and incomplete lineage sorting. This type of data represents the history of a single gene that might not be representative of a species history, which can lead to ambiguities in assessing species boundaries (Hickerson *et al.*, 2006; Knowles & Carstens, 2007; Leliaert *et al.*, 2014). More recently, multi-locus genetic data has provided potential evidence of divergence even at an early stage of diversification (Knowles & Carstens, 2007). Methods for multi-marker species delimitation have been developed mainly as a response to technical advances in sequencing technology that allow for the generation of large-scales datasets (Faircloth *et al.*, 2012; Lemmon *et al.*, 2012; Lemmon & Lemmon, 2013). The advent of Massively Parallel Sequencing (MPS) and the development of generalized data collection facilitated the acquisition of multi-locus datasets for large numbers of individuals and have been an important tool to resolve the interrelationships of longstanding problematic taxa (Crawford *et al.*, 2012; Faircloth *et al.*, 2012; McCormack *et al.*, 2013).

Ultraconserved Elements (UCEs) markers have been recently developed and can be used in conjunction with sequence capture and MPS to generate large amounts of orthologous sequence data among a taxonomically diverse set of species (Faircloth *et al.*, 2012). UCEs were originally discovered in mammals (Bejerano *et al.*, 2004) but are common in a diversity of

Arthropods (Blaimer *et al.*, 2016; Faircloth *et al.*, 2015; Longino & Branstetter, 2020). These elements are short stretches of highly conserved DNA shared across distantly related taxa, which harbor variation in its flanking regions. The variation in the flanks increases with distance from the core UCE, providing information for estimating phylogenies at multiple evolutionary timescales (Smith *et al.*, 2014; Crawford *et al.*, 2015; Faircloth *et al.*, 2012). Because of the increase in variation in the flanking regions of UCEs, these markers are suitable for studies at narrow timescales. One of the advantages of using UCEs is that they work with suboptimally preserved specimens and/or degraded DNA (McCormack *et al.*, 2015; Blaimer *et al.*, 2016).

With multilocus genetic data, coalescent models have been implemented to identify the evolutionary processes that contribute to speciation (Fujita *et al.*, 2012). Coalescence-based methods highlight incomplete lineage-sorting as a source of inconsistency between gene trees and the species trees (Rannala & Yang, 2003; Liu *et al.*, 2009; Heled & Drummond, 2010), inferring species limits based on a rigorous population framework (Fujita *et al.*, 2012; Rannala, 2015). With the addition of data sources, coalescent-based species delimitation might play an important role in an integrative taxonomy that emphasizes the identification of species and the processes that have promoted lineage diversification (Fujita *et al.*, 2012). The development of species delimitation methods that are based on the multispecies coalescent model (MSC) was an important advance in biodiversity research (Takahata *et al.*, 1995; Rannala & Yang, 2003). Several studies have delimited species using this approach (e. g. Carstens & Dewey, 2010; Leaché & Fujita, 2010; Carstens & Satler, 2013).

Morphometric and molecular approaches have been successfully applied in studies of bees' biogeography, systematic and evolution (Quezada-Euán *et al.*, 2007; Quezada-Euán *et al.*, 2015; Santos *et al.*, 2019). However, there are relatively few studies using integrative taxonomy to assess the relationships and species thresholds in euglossine bees. Orchid bees (Apidae: Euglossini) are endemic bees of the Neotropical Region, so-called because of the strict relationship of males with most orchids in this region, and are also known because of their brilliant metallic integumental coloration (Michener, 2007). Males of orchid bees collect aromatic substances found in several plant species, especially orchids, and nonfloral resources (Dressler, 1982) to use these substances principally as a signal to attract females and communicate availability for mating (Eltz *et al.*, 2011). Five Euglossini genera are recognized: *Euglossa* Latreille 1802, *Eufrisea* Cockerell 1908, *Eulaema* Lepelletier 1841, *Exaerete* Hoffmannsegg 1817, and *Aglae* Lepelletier & Serville 1825, encompassing approximately 250 species (Michener, 2007; Moure *et al.*, 2012). *Eulaema* contains the largest species of orchid bees, with body size varying from 18 to 30mm in length (Oliveira, 2000; Melo, 2014), and do

not exhibit a metallic shine in the head and thorax as many other Euglossini do (Oliveira, 2000).

The species *Eulaema (Apeulaema) cingulata* (Fabricius, 1804) has a wide distribution associated mainly to forested areas in the Neotropical region, being frequently sampled in faunistic inventories and widely used for ecological studies (e.g. Cavalcante *et al.*, 2012, Rocha-Filho & Garófalo, 2015, Marques *et al.*, 2017). However, this name is involved in conflicting hypotheses of interspecific delimitation with *Eulaema (Apeulaema) pseudocingulata* Oliveira, 2006. This second nominal species was described based on few morphological differences from *E. cingulata* and due to its geographic distribution, apparently restricted to the Amazon Forest (Oliveira, 2006). The taxonomic status of *Eulaema cingulata* and *E. pseudocingulata* has been subject of debate. Males of both species can only be separated phenotypically by the velvety area on the middle leg, which is much narrower and far from the posterior edge in *Eulaema pseudocingulata*, whereas in *E. cingulata* it is wide and close to the posterior edge. In addition, the coloration of the male abdomen is relatively darker in *E. pseudocingulata* (Figure 1), which description was carried out based only on morphological aspects of males as the female is unknown. In the present study we evaluate the differences between specimens identified as *Eulaema cingulata* and *Eulaema pseudocingulata*, employing an integrative taxonomic approach that combines morphological and molecular data in order to test species identities. Our expectation was to clarify objectively whether different morphs correspond to one or two evolutionary units.



Figure 1. A-B. Male specimens of *Eulaema cingulata* from Sergipe, Brazil. (A) Body in dorsal view. (B) Velvety area in the mid tibiae. C-D. Male specimens of *E. pseudocingulata* from Pará, Brazil. (C) Body in dorsal view. (D) Velvety area in the mid tibiae.

2. MATERIAL AND METHODS

2.1 Sampling and DNA extraction

The material selected for the present study covers most of the reported distribution of *Eulaema cingulata* and *E. pseudocingulata*, especially in the areas they appear to be sympatric (Amazon Forest). A total of 107 male specimens were used for the Geometric Morphometrics analysis, being 53 and 54 specimens of *E. pseudocingulata* and *E. cingulata*, respectively. For molecular analyses we obtained DNA sequences from 19 recently-collected males, maximum age of 10 years after collection, used in COI (n= 8) and UCE (n= 15) analyses, while some samples were used in both molecular approaches (n= 4). Four sequences of the mitochondrial gene Cytochrome oxidase I (COI) were obtained from GenBank and two sequences were obtained from BOLD Systems. We added five other *Eulaema* species as outgroups in our molecular analyses: *E. mocsaryi* and *E. nigrita* for COI analysis, and *E. bombiformis*, *E. chocoana*, *E. meriana*, *E. mocsaryi* and *E. nigrita* for the phylogenomic analysis. Information of the specimen included, and their repositories, are presented in Table 1.

We extracted total genomic DNA from the hind legs or thoracic muscles of the specimens using Qiagen DNeasy Kits (Qiagen™), including modifications from Evangelista (2012). Some specimens used were directly preserved in 100% ethanol right after collecting, whereas others were kept dried mounted for various periods of time. Before extraction, pinned specimens were left in a humid chamber for 24 hours in order to facilitate the removal of the leg or abdomen muscle without damaging the whole specimen. Then the samples were digested with buffer plus 20µL of protease K and after six hours we added more 10µL. The incubation was performed overnight (up to 12 hours).

Table 1. Voucher ID, locality data and institutional repositories of the specimens used in the present study. INPA = Instituto Nacional de Pesquisas da Amazônia; MZUSP = Museu de Zoologia da Universidade de São Paulo; UENF = Universidade Estadual do Norte Fluminense Darcy Ribeiro; UFAC = Universidade Federal do Acre; UFMG = Universidade Federal de Minas Gerais.

Species	Country	State/ Province	Latitude	Longitude	Date	Geometric morphometrics		COI	UCEs	Repository
						Wings	Head			
<i>E. bombiformis</i>	Ecuador	Orellana	0°27'16.26"S	76°59'42.10"W	02.viii.2009	-	-	-	ML256	López-Uribe lab.
<i>E. chocoana</i>	Colômbia	Chocó	5°29'58.86"N	76°32'33.35"W	18.vi.2003	-	-	-	ML023	López-Uribe lab.
<i>E. cingulata</i>	Brazil	Acre	09°59'14.3"S	67°50'29.0"W	25.i.2016	EC011	-	-	-	UFAC
<i>E. cingulata</i>	Brazil	Acre	07°26'45.20"S	72°54'37.20"W	24.vi.2017	EC010	-	-	-	UFAC
<i>E. cingulata</i>	Brazil	Alagoas	09°12'42.1"S	35°52'12.0"W	13.v.2015	EC017	EC017	-	-	UFMG
<i>E. cingulata</i>	Brazil	Alagoas	09°12'42.1"S	35°52'12.0"W	13.v.2015	EC016	EC016	-	-	UFMG
<i>E. cingulata</i>	Brazil	Alagoas	09°12'42.1"S	35°52'12.0"W	13.v.2015	EC015	EC015	-	-	UFMG
<i>E. cingulata</i>	Brazil	Amazonas	-	-	-	EC043	-	-	-	INPA
<i>E. cingulata</i>	Brazil	Amazonas	-	-	ix.1992	EC039	-	-	-	INPA
<i>E. cingulata</i>	Brazil	Amazonas	03°56'02"S	61°19'03"W	29.xi- 08.xii.2003	EC042	-	-	-	INPA
<i>E. cingulata</i>	Brazil	Amazonas	4°54'45.30"S	61°06'30.66"W	19.vii.2007	EC052	-	-	-	INPA
<i>E. cingulata</i>	Brazil	Amazonas	2°32'00.73"S	57°45'22.08"W	02.iv.2011	EC030	EC030	-	-	UFMG
<i>E. cingulata</i>	Brazil	Amazonas	2°32'00.73"S	57°45'22.08"W	02.iv.2011	-	AM02	-	-	UFMG
<i>E. cingulata</i>	Brazil	Amazonas	07°21'46.7"S	71°52'07.1"W	11-15.v.2011	EC037	-	-	-	INPA
<i>E. cingulata</i>	Brazil	Amazonas	3°07'11.69"S	60°01'14.22"W	22.vii.2012	EC045	-	-	-	INPA
<i>E. cingulata</i>	Brazil	Amazonas	6°30'24.84"S	64°33'05.4"W	07.v.2013	EC044	-	-	-	INPA
<i>E. cingulata</i>	Brazil	Amazonas	4°05'35.47"S	63°08'25.05"W	16-17.vi.2017	EC046	-	-	-	INPA
<i>E. cingulata</i>	Brazil	Bahia	15°47'45.7"S	40°31'57.0"W	30.i.2014	EC033	EC033	-	-	UFMG
<i>E. cingulata</i>	Brazil	Bahia	17°06'18.6"S	39°20'18.1"W	12.xii.2015	EC032	EC032	-	-	UFMG
<i>E. cingulata</i>	Brazil	Bahia	16°52'48.9"S	39°24'48.2"W	14.xii.2015	EC031	EC031	-	-	UFMG
<i>E. cingulata</i>	Brazil	Espírito Santo	19°06'- 19°18"S	39°45'-40°19'W	24.iv.1997	EC049	-	-	-	INPA

Table 1. Voucher ID, locality data and institutional repositories of the specimens used in the present study (continued).

<i>E. cingulata</i>	Brazil	Espírito Santo	18°23.28,1"S	39°50'92.0"W	5.vii.2014	EC029	EC029	-	-	UFMG
<i>E. cingulata</i>	Brazil	Espírito Santo	19°03'05.4"S	40°04'15.3"W	17.xii.2015	EC025	EC025	-	-	UFMG
<i>E. cingulata</i>	Brazil	Espírito Santo	19°00'30.45"S	40°06'37.65"W	26-27.v.2015	EC024	EC024	-	-	UFMG
<i>E. cingulata</i>	Brazil	Maranhão	3°47'15.52"S	46°21'44.89"W	30.vi.2010	EC041	-	-	-	INPA
<i>E. cingulata</i>	Brazil	Mato Grosso	15°49'30.82"S	55°24'30.10"W	6.xii.2012	EC026	EC026	-	-	UFMG
<i>E. cingulata</i>	Brazil	Mato Grosso	15°49'30.82"S	55°24'30.10"W	6.xii.2012	EC027	-	-	-	UFMG
<i>E. cingulata</i>	Brazil	Mato Grosso	4°57'16.2"S	59°01'39.6"W	21.vii.2014	-	-	TA07	TA07	UFMG
<i>E. cingulata</i>	Brazil	Minas Gerais	19°40'07.96"S	42°31'52.84"W	24.ix.1989	EC038	-	-	-	INPA
<i>E. cingulata</i>	Brazil	Minas Gerais	19°40'07.96"S	42°31'52.84"W	26.iv.1990	EC050	-	-	-	INPA
<i>E. cingulata</i>	Brazil	Minas Gerais	19°14'53.4"S	43°18'14.5"W	8.x.2014	EC012	EC012	-	-	UFMG
<i>E. cingulata</i>	Brazil	Minas Gerais	19°04'08.1"S	43°13'14.3"W	21.i.2015	EC013	EC013	-	-	UFMG
<i>E. cingulata</i>	Brazil	Minas Gerais	20°06'02.7"S	44°18'11.6"W	3.xi.2015	EC014	EC014	-	-	UFMG
<i>E. cingulata</i>	Brazil	Pará	6°22'29"S	50°22'56"W	27.iv.2010	EC020	EC020	-	-	UFMG
<i>E. cingulata</i>	Brazil	Pará	-	-	07-09.vii.2012	EC006	-	-	-	MZUSP
<i>E. cingulata</i>	Brazil	Pará	-	-	07-09.vii.2012	EC007	-	-	-	MZUSP
<i>E. cingulata</i>	Brazil	Pará	-	-	07-09.vii.2012	EC008	-	-	-	MZUSP
<i>E. cingulata</i>	Brazil	Pará	06°41'37.2"S	50°31'05.0"W	3.iv.2015	EC021	EC021	-	-	UFMG
<i>E. cingulata</i>	Brazil	Pará	06°24'10.9"S	49°58'26.3"W	12.ii.2015	-	-	TA08	TA08	UFMG
<i>E. cingulata</i>	Brazil	Paraíba	06°44'20.6"S	35°10'17.8"W	26.vii.2014	-	PB02	-	-	UFMG

Table 1. Voucher ID, locality data and institutional repositories of the specimens used in the present study (continued).

<i>E. cingulata</i>	Brazil	Paraíba	06°42'48.5"S	35°11'46.8"W	27.vii.2014	EC028	EC028	-	-	UFMG
<i>E. cingulata</i>	Brazil	Paraíba	06°43'18.9"S	35°12'13.3"W	27.vii.2014	-	-	-	TA09	UFMG
<i>E. cingulata</i>	Brazil	Rio de Janeiro	21°47'23"S	41°26'15"W	26.ix.2008	-	RJ02	-	-	UENF
<i>E. cingulata</i>	Brazil	Rio de Janeiro	21°47'23"S	41°26'15"W	26.ix.2008	-	RJ02	-	-	UENF
<i>E. cingulata</i>	Brazil	Rio de Janeiro	22°37'16.9"S	43°47'05.0"W	24.ii.2012	-	RJ01	-	-	UFMG
<i>E. cingulata</i>	Brazil	Rio de Janeiro	22°28'47.27"S	42°12'10.45"W	25.v.2013	-	RJ03	-	-	UENF
<i>E. cingulata</i>	Brazil	Rondônia	9°35'24.00"S	65°03'00.00"W	5.iii.2010	EC002	-	-	-	MZUSP
<i>E. cingulata</i>	Brazil	Rondônia	-	-	8.iii.2010	-	EC001	-	TA12	MZUSP
<i>E. cingulata</i>	Brazil	Rondônia	9°36'00.00"S	65°21'36.00"W	10.iii.2010	EC003	-	-	-	MZUSP
<i>E. cingulata</i>	Brazil	Rondônia	12°13'56.6"S	61°24'24.9"W	6.xii.2012	EC019	EC019	-	-	UFMG
<i>E. cingulata</i>	Brazil	Rondônia	12°13'56.6"S	61°24'24.9"W	7.xii.2012	EC018	EC018	-	-	UFMG
<i>E. cingulata</i>	Brazil	Rondônia	9°7'30"S	62°55'0"W	08.xi.2013	EC051	-	-	-	INPA
<i>E. cingulata</i>	Brazil	Roraima	3°45'19.52"N	61°43'03.22"W	14.vii.2009	EC047	-	-	-	INPA
<i>E. cingulata</i>	Brazil	Roraima	-	-	18-21.xii.2017	EC040	-	-	-	INPA
<i>E. cingulata</i>	Brazil	Sergipe	10°46'02.3"S	37°20'32.9"W	21.vii.2014	EC035	EC035	-	-	UFMG
<i>E. cingulata</i>	Brazil	Sergipe	10°45'26.7"S	37°20'53.5"W	22.vii.2014	EC034	EC034	-	-	UFMG
<i>E. cingulata</i>	Brazil	Tocantins	10°08'S	48°20'W	3.i.2015	EC036	-	-	-	INPA
<i>E. cingulata</i>	Colômbia	Medellin	6°13'31.44"N	75°32'13.46"W	i.2014	-	-	TA10	TA10	UFMG
<i>E. cingulata</i>	Colômbia	Monteria	8°42'54.96"N	75°51'48.21"W	i.2014	CO01	CO01	-	-	UFMG
<i>E. cingulata</i>	Costa Rica	Cartago	9°54'24.14"N	83°40'48.18"W	11.vi.1976	EC048	-	-	-	INPA
<i>E. meriana</i>	Costa Rica	La Selva	9°33'25.93"N	82°55'55.93"W	06.vi.2008	-	-	-	ML089	López-Uribe lab.
<i>E. meriana</i>	Peru	Madre de Dios	12°28'50.11"S	70°20'32.21"W	3.xi.1997	-	-	MN345153.1	-	GenBank

Table 1. Voucher ID, locality data and institutional repositories of the specimens used in the present study (continued).

<i>E. mocsaryi</i>	Colômbia	Puerto Carreño	6°11'23.68"N	67°28'57.25"W	2.x.2002	-	-	EU421502	-	GenBank
<i>E. nigrita</i>	Brazil	Rondônia	09°26'08.2"S	64°48'09.6"W	02-16.x.2013	-	-	TA25	-	
<i>E. pseudocingulata</i>	Brazil	Acre	9°58'31.21"S	68°25'45.56"W	03.ii.1997	EC022	EC022	-	-	UFMG
<i>E. pseudocingulata</i>	Brazil	Acre	9°58'31.21"S	68°25'45.56"W	03.ii.1997	EC023	EC023	-	-	UFMG
<i>E. pseudocingulata</i>	Brazil	Acre	9°58'31.21"S	68°25'45.56"W	03.ii.1997	EP029	EP029	-	-	UFMG
<i>E. pseudocingulata</i>	Brazil	Acre	9°58'31.21"S	68°25'45.56"W	05.x.1997	EP028	EP028	-	-	UFMG
<i>E. pseudocingulata</i>	Brazil	Acre	10°02'01.5"S	67°46'14.2"W	15.ix.2015	EP021	-	-	-	UFAC
<i>E. pseudocingulata</i>	Brazil	Acre	09°56'47.3"S	67°49'41.1"W	18.xi.2015	EP008	-	-	-	UFAC
<i>E. pseudocingulata</i>	Brazil	Acre	10°02'01.5"S	67°46'14.2"W	24.xi.2015	EP039	-	-	-	UFAC
<i>E. pseudocingulata</i>	Brazil	Acre	10°03'15.6"S	67°49'30.8"W	14.xii.2015	EP015	EP015	-	-	UFAC
<i>E. pseudocingulata</i>	Brazil	Acre	10°02'01.5"S	67°46'14.2"W	17.i.2016	EP016	EP016	-	-	UFAC
<i>E. pseudocingulata</i>	Brazil	Acre	10°02'01.5"S	67°46'14.2"W	17.i.2016	-	-	TA01	-	UFAC
<i>E. pseudocingulata</i>	Brazil	Acre	09°59'14.3"S	67°50'29.0"W	25.i.2016	EP007	-	-	-	UFAC
<i>E. pseudocingulata</i>	Brazil	Acre	09°57'14.0"S	67°47'34.7"W	26.i.2016	EP033	EP033	-	-	UFAC
<i>E. pseudocingulata</i>	Brazil	Acre	10°03'15.6"S	67°49'30.8"W	02.ii.2016	EP013	EP013	-	-	UFAC
<i>E. pseudocingulata</i>	Brazil	Acre	09°57'11.8"S	67°47'33.4"W	7.ii.2016	EC009	-	TA06	TA06	UFAC
<i>E. pseudocingulata</i>	Brazil	Acre	10°02'20.9"S	67°47'38.6"W	15.xi.2016	EP009	-	-	-	UFAC
<i>E. pseudocingulata</i>	Brazil	Acre	07°28'37.0"S	072°54'06.5"W	12.i.2017	EP014	EP014	-	-	UFAC
<i>E. pseudocingulata</i>	Brazil	Amapá	0°01'44.25"N	51°04'03.96"W	iii.1991	EP050	-	-	-	INPA
<i>E. pseudocingulata</i>	Brazil	Amapá	0°2'56.35"S	51°7'45.45"W	31.iii.2012	EP043	-	-	-	INPA
<i>E. pseudocingulata</i>	Brazil	Amazonas	05°15'39"S	60°42'32"W	23.iv.2005	EP041	-	-	-	INPA
<i>E. pseudocingulata</i>	Brazil	Amazonas	07°19'10"S	64°40'07"W	13.vi- 07.vii.2006	EP044	-	-	-	INPA
<i>E. pseudocingulata</i>	Brazil	Amazonas	00°25'09.3"N	63°23'00.7"W	30.vii.2007	EP040	-	-	-	INPA
<i>E. pseudocingulata</i>	Brazil	Amazonas	2°57'48.04"S	59°55'22.20"W	14.viii.2008	EP042	-	-	-	INPA
<i>E. pseudocingulata</i>	Brazil	Goiás	18°15'38"S	52°53'02"W	12.v.2014	-	-	TA22	-	UFMG
<i>E. pseudocingulata</i>	Brazil	Goiás	18°15'38"S	52°53'02"W	4.x.2014	EP025	EP025	-	-	UFMG

Table 1. Voucher ID, locality data and institutional repositories of the specimens used in the present study (continued).

<i>E. pseudocingulata</i>	Brazil	Goiás	18°15'38"S	52°53'02"W	4.x.2014	-	-	-	TA02	UFMG
<i>E. pseudocingulata</i>	Brazil	Maranhão	6°37'56.29"S	45°53'4.25"W	18.iii.2013	EP022	EP022	-	-	UFMG
<i>E. pseudocingulata</i>	Brazil	Mato Grosso	13°49'20.49"S	56°05'00.76"W	14.i.2000	EP046	-	-	-	INPA
<i>E. pseudocingulata</i>	Brazil	Mato Grosso	17°18'56.77"S	53°13'06.97"W	03.iii.2010	EP030	EP030	-	-	UFMG
<i>E. pseudocingulata</i>	Brazil	Mato Grosso	13°48'55.5"S	59°41'11.1"W	18.vii.2014	EP024	EP024	-	-	UFMG
<i>E. pseudocingulata</i>	Brazil	Mato Grosso	14°57'16.1"S	59°01'32.4"W	21.vii.2014	EP023	EP023	-	-	UFMG
<i>E. pseudocingulata</i>	Brazil	Mato Grosso	14°57'16.2"S	59°01'39.6"W	21.vii.2014	-	-	-	TA03	UFMG
<i>E. pseudocingulata</i>	Brazil	Pará	1°27'20.72"S	48°29'24.65"W	iii.1954	EP005	-	-	-	MZUSP
<i>E. pseudocingulata</i>	Brazil	Pará	8°56'56.9"S	54°54'05.62"W	viii.1995	EP004	-	-	-	MZUSP
<i>E. pseudocingulata</i>	Brazil	Pará	1°28'01.09"S	56°22'45.39"W	12.xii.2006	EP020	EP020	-	-	UFMG
<i>E. pseudocingulata</i>	Brazil	Pará	1°28'01.09"S	56°22'45.39"W	24.ii.2007	EP035	EP035	-	-	UFMG
<i>E. pseudocingulata</i>	Brazil	Pará	5°35'33.76"S	50°10'59.70"W	03.viii.2010	EP038	EP038	-	-	UFMG
<i>E. pseudocingulata</i>	Brazil	Pará	5°35'33.76"S	50°10'59.70"W	03.viii.2010	EP034	EP034	-	-	UFMG
<i>E. pseudocingulata</i>	Brazil	Pará	06°24'09"S	50°20'43"W	25.iv.2010	EP036	EP036	-	-	UFMG
<i>E. pseudocingulata</i>	Brazil	Pará	6°23'01"S	50°21'28"W	9.ix.2010	-	-	-	TA04	UFMG
<i>E. pseudocingulata</i>	Brazil	Pará	5°35'33.76"S	50°10'59.70"W	12.ii.2011	EP032	EP032	-	-	UFMG
<i>E. pseudocingulata</i>	Brazil	Pará	1°45'42.59"S	55°51'49.74"W	31.iii.2011	-	PA08	-	-	UFMG
<i>E. pseudocingulata</i>	Brazil	Pará	03°31'40"S	51°52'56"W	03-05.vii.2012	EP006	-	-	-	MZUSP
<i>E. pseudocingulata</i>	Brazil	Pará	-	-	17.vi.2017	EP012	EP012	-	-	MPEG
<i>E. pseudocingulata</i>	Brazil	Pará	-	-	09.vii.2017	EP010	EP010	-	-	MPEG
<i>E. pseudocingulata</i>	Brazil	Pará	07°55'33"S	49°24'11"W	14.vii.2017	EP011	EP011	TA19	-	MPEG
<i>E. pseudocingulata</i>	Brazil	Rondônia	9°25'12.00"S	64°49'12.00"W	3.iii.2010	EP001	-	-	-	MZUSP
<i>E. pseudocingulata</i>	Brazil	Rondônia	9°35'24.00"S	65°03'00.00"W	3.iii.2010	EP002	-	-	-	MZUSP
<i>E. pseudocingulata</i>	Brazil	Rondônia	9°36'00.00"S	65°21'36.00"W	10.iii.2010	EP003	-	-	-	MZUSP
<i>E. pseudocingulata</i>	Brazil	Rondônia	8°45'40.30"S	63°54'01.58"W	21.xi.2010	-	-	-	TA05	MZUSP

Table 1. Voucher ID, locality data and institutional repositories of the specimens used in the present study (continued).

<i>E. pseudocingulata</i>	Brazil	Rondônia	8°49'33.94"S	63°57'47.16"W	22.xi.2011	EP026	EP026	-	-	UFMG
<i>E. pseudocingulata</i>	Brazil	Rondônia	8°49'33.94"S	63°57'47.16"W	10.i.2012	EP027	EP027	-	-	UFMG
<i>E. pseudocingulata</i>	Brazil	Rondônia	12°13'56.6"S	61°24'24.9"W	9.xii.2012	EP037	-	-	-	UFMG
<i>E. pseudocingulata</i>	Brazil	Roraima	3°25'00.00"N	61°40'00.00"W	20-30.iii.1987	EP045	-	-	-	INPA
<i>E. pseudocingulata</i>	Brazil	Roraima	3°25'00.00"N	61°40'00.00"W	5-10.x.1987	EP047	-	-	-	INPA
<i>E. pseudocingulata</i>	Brazil	Roraima	3°45'19.52"N	61°43'03.22"W	14.vii.2009	EP048	-	-	-	INPA
<i>E. pseudocingulata</i>	Brazil	Roraima	03°46'56.2"N	61°43'15.3"W	15.vii.2014	EP017	EP017	-	-	UFMG
<i>E. pseudocingulata</i>	Brazil	Roraima	03°46'56.2"N	61°43'15.3"W	15.vii.2014	EP018	EP018	-	-	UFMG
<i>E. pseudocingulata</i>	Brazil	Roraima	02°52'38.4"N	60°43'13.1"W	24.vii.2014	EP019	EP019	-	-	UFMG
<i>E. pseudocingulata</i>	Brazil	Roraima	02°52'38.4"N	60°43'13.1"W	24.vii.2014	EP031	EP031	-	-	UFMG
<i>E. pseudocingulata</i>	Brazil	Roraima	2°44'47.93"N	62°11'46.40"W	18-21.xii.2017	EP049	-	-	-	INPA
<i>E. pseudocingulata</i>	French Guiana	Cayenne	4°55'20.71"N	52°18'48.43"W	1.i.2004	-	-	DWR0071	-	BoldSystems
<i>E. pseudocingulata</i>	Peru	Madre de Dios	12°28'50.11"S	70°20'32.21"W	23.vii.2010	-	-	-	ML510	López-Urbe lab.
<i>E. sororia</i>	Colômbia	Chocó	5°29'58.86"N	76°32'33.35"W	30.vii.2006	-	-	-	ML029	López-Urbe lab.

2.2. Morphometric data

We use two different approaches for the geometric morphometric analysis, one considering the species and the other considering the geographical origins (Amazon Forest and Atlantic Forest). The specimen from the State of Goiás (Brazil) was removed in the last analysis, since it was collected in a Savannah area. Specimens were photographed using a Leica DFC 295® camera attached to a stereomicroscope Leica M205C®, taking all precautions to avoid possible distortions. From each specimen the right forewing was separated from the body at the base of the radial vein with the help of forceps and fixed between glass microscopy slides to keep it flat. The heads were always photographed in frontal view. The photographs were transformed into TPS files using the software TpsUtil 1.60 (Rohlf, 2013). Landmarks were set using the software tpsDig version 2.26 (Rohlf, 2006). We selected 18 homologous landmarks on the vein intersections for the wing analysis and 10 landmarks for the head (Figure 2).

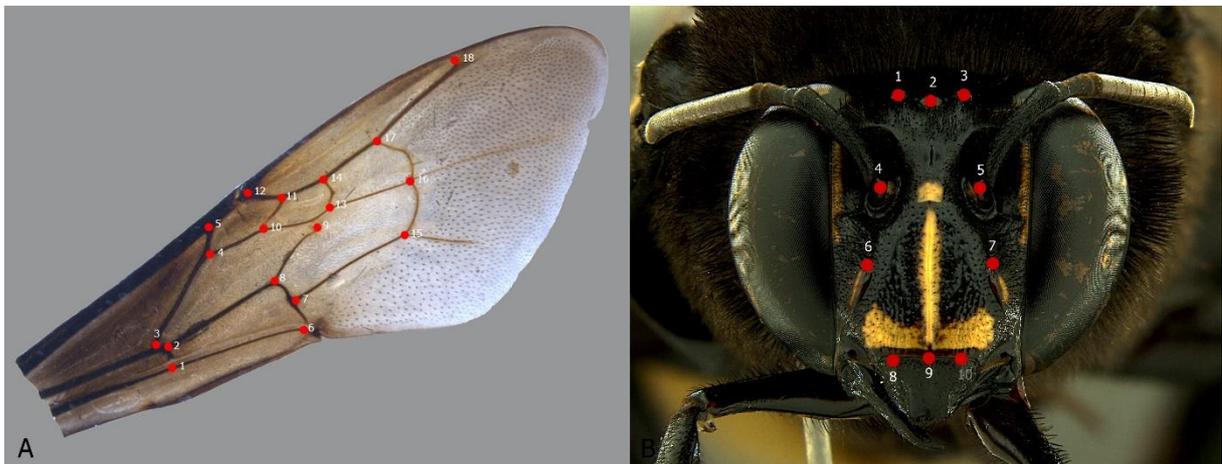


Figure 2. Landmarks used in the geometric morphometric analyses of (A) forewings and (B) head in frontal view of males of *Eulaema cingulata* and *Eulaema pseudocingulata*.

For comparison of overall wing and head sizes between species and geographical groups, we extracted the centroid size information obtained from the Procrustes Analysis. A regression analysis was performed to quantify the effect of allometry in our data. Afterwards, we removed this allometric effect in order to quantify independently the shape variation. Shape variables were obtained by performing a generalized Procrustes analysis which eliminates the effects of orientation, position and size by scaling all configurations to the same centroid size (Klingenberg, 2015). The resulting landmark configurations retain only shape information without effects of position, orientation and size. A Principal Component Analysis (PCA) was

carried out using the relative Cartesian coordinates of each landmark after alignment. The shape difference between species and geographical clusters were tested using a Discriminant Analysis followed by a leave-one-out Cross-validation test (Lachenbruch, 1967). All analyses were performed using the software MorphoJ (Klingenberg, 2011). Additionally, we used the percentages of correct classification to evaluate the discriminatory power of wing and head shapes.

2.3. Amplification and sequencing of COI

Mitochondrial cytochrome oxidase I (COI) was amplified using the commonly employed barcode fragment (~700 bp) for species delimitation: primers LCO (5'-GGTCAACAAATCATAAAGATATTGG-3') and HCO (5'-TAAACTTCAGGGTGACCAAAAAATCA-3') (Folmer *et al.*, 1994). Polymerase Chain Reaction (PCR) amplifications were performed in a final volume of 17µL including 0.13 µL of Taq Polymerase, 2µL of genomic template, 1.3µL of each primer, 3.4µL of dNTPs (10mM), 0.85µL of MgCl, 1.7µL of 10x Qiagen Buffer; the remaining volume (6.32µL) was filled with purified water. Amplification were performed with an initial step of three minutes at 94°C, followed by 40 cycles of 30 seconds at 94°C, 30 seconds for annealing at 50°C and 1 min at 72 °C. After 40 cycles, a final step was performed at 72°C for 10 minutes.

All steps related to DNA extraction and amplification were performed at the Molecular Biology Laboratory of the Museu de Zoologia da Universidade de São Paulo, Brazil (MZSP). Successful amplifications were confirmed with gel electrophoresis and sent to Macrogen (Seoul, South Korea) for post-PCR clean-up and Sanger sequencing. PCR products were sequenced in two directions and sequence quality was determined by the quality scores provided by Macrogen and by visually examining the chromatograms using the software Ugene (Okonechnikov *et al.*, 2012).

2.4. UCE library preparation

We used the Ultra Conserved Element (UCE) approach to phylogenomics (Faircloth *et al.*, 2012) to generate a genome-scale data set. This method combines the targeted enrichment of thousands of UCE loci with multiplexed next-generation sequencing. A recent published bait set specific to bees, ants and other apoid wasps was used to perform the enrichment (“hym-v2-bee-ant-specific”; Grab *et al.*, 2019). This bait set is a subset of the principal Hymenoptera bait set first reported in Branstetter *et al.* (2017). Each DNA sample was sheared using a Qsonica Q800R2 acoustic sonicator, with the target fragment size range being 400-600 bp (60–120 secs

shear time, 25% amplitude, 10–10 sec pulse). Having larger fragment sizes in the sequencing pool improves the amount of flanking DNA that gets sequenced and improves assembled contig lengths. For recently collected, high quality samples, the sonicator was run for 90–120 seconds (shearing time) at 25% amplitude and with a 10–10 second on-off pulse. For relatively older samples that likely had more degraded DNA, we adjusted the shearing times to between 30–60 seconds. Following sonication, fragmented DNA was cleaned at 3x volume using a homemade SPRI-bead substitute (Rohland & Reich, 2012). Illumina sequencing libraries were generated for each sample using Kapa Hyper Prep Kits and custom, dual-indexing adapters (Glenn *et al.*, 2016). Libraries were PCR amplified for 12 cycles for most samples, but for lower quality samples we increased the number to 14–16 cycles. Each amplified library was cleaned using 1.0–1.2x SPRI beads in order to remove contaminants and to select out small fragments below 200 bp, including adapter dimer. The concentration of the final cleaned library was measured using Qubit.

Libraries were PCR amplified for 12 cycles, purified using an Ampure XP substitute (Rohland & Reich, 2012), and quantified using Qubit. To enrich UCE loci, 10 samples were pooled at equimolar concentrations and then up to 500 ng of each pool was enriched following the manufacturer’s protocol for day 1 (MYcroarray enrichment protocol v3.02) and the standard UCE protocol for day 2 (enrichment protocol v1.5 available at ultraconserved.org). For each enrichment, the custom bait set was diluted 1:4 (1 μ L bait, 4 μ L H₂O) and we performed the enrichment incubation at 65°C for 24 hours using strip tubes and a PCR thermal cycler. For the second day of enrichment, we used 50 μ L of streptavidin beads per sample and performed on-bead PCR following the three heated (65°C) wash steps. The enriched pools were amplified for 18 cycles and the resulting products were cleaned with SPRI beads at 1x volume. Following enrichment, each enrichment pool was quantified using qPCR and pooled together into a final sequencing pool at equimolar concentrations. Sequencing pools were sent to either the University of Utah Genomics Core for sequencing on an Illumina HiSeq 2500 (2x125, v4 chemistry), or to Novogene for sequencing on HiSeq X.

2.5. Bioinformatics

COI sequences were aligned using multiple sequence alignment online tool MAFFT (Kato *et al.*, 2019) and manually edited using the software Ugene (Okonechnikov *et al.*, 2012). For the phylogenomic analysis, we cleaned the reads for adapter contamination and low-quality bases using Illumiprocessor (Faircloth, 2013) incorporated in the software Trimmomatic (Bolger *et al.*, 2014). We assembled reads *de novo* for each individual using Spades (Bankevich

et al., 2012). To identify UCE regions from the bulk of assembled contigs and to remove paralogs we employed the HymV2-bee-ant UCE bait files from Branstetter (2017). We aligned individually all the loci using MAFFT (Katoh & Standley, 2013), and we trimmed resulting alignments using GBLOCKS (Castresana, 2000; Tavalera & Castresana, 2007). We removed loci that had data for fewer than 75% of taxa and generated a concatenated matrix from the resulting alignment set.

2.6. Molecular approaches for species phylogeny and delimitation

Estimation of the genetic distance among specimens was calculated by MEGA-X (Kumar *et al.*, 2018) with 10,000 bootstraps. We inferred phylogenetic trees using Maximum Likelihood (ML), Bayesian Inference (BI) and Species Tree (ST) methods. Maximum likelihood analyses were inferred using two different strategies: single concatenated alignment and partitioned based on the best-fitting partitioning scheme. We test these two phylogenetic model-based methods since concatenated-based methods infer phylogenies that assume that all genes have similar histories; on the other hand, when the dataset is partitioned, different models of evolution are applied to individual loci (Young & Gillung, 2020). The best-fitting partitioning scheme was obtained using Sliding-Window Site Characteristics (SWSC). This approach divides each UCE into three data blocks, corresponding to the right flank, core and left flank (Tagliacolo & Lanfear, 2018). After running the SWSC-EN algorithm, the resulting data subsets were analyzed using PartitionFinder2 (Lanfear *et al.*, 2017). For this analysis, we used the *reclusterf* algorithm, AICc model selection criterion and GTR+G model of sequence evolution. The best partitioning scheme grouped loci having the same substitution model to be used in the subsequent analyses. We used the likelihood-based program Iq-Tree v.1.6.12 (Nguyen *et al.*, 2015) for both partitioning schemes. The substitution model of the concatenated alignment (TVM+F+R2) was obtained with the ModelFinder implemented in the Iq-Tree software (Kalyaanamoorthy *et al.*, 2017). To assess branch support, we performed 1,000 replicates of the ultrafast bootstrap approximation (UFBoot; Hoang *et al.*, 2018). Additionally, 1,000 replicates of the branch-based Shimodaira-Hasegawa approximate likelihood-rate test (SH-aLRT; Guindon *et al.*, 2010) were conducted using the command ‘-alrt’. Only clades with support values of UFBoot ≥ 0.95 and SH-aLRT ≥ 0.80 were considered robust.

We performed Bayesian phylogenetic analyses in MrBayes v.3.2 (Ronquist *et al.*, 2012) using the best nucleotide model estimated by PartitionFinder2 (Lanfear *et al.*, 2017). The Markov Chain Monte Carlo (MCMC) was run for 20 million generations sampled every 1000th generation. Twenty-five percent of the first trees were discarded as burnin. Bayesian trees were

visualized and edited in FigTree v. 1.4.4 (Rambaut, 2016).

To account for heterogeneous gene histories that may influence phylogenetic accurate resolution, we inferred a species trees under the multi-species coalescent model using the program ASTRAL-III v.5.7.3 (Zhang *et al.*, 2018). For this analysis, we used the set of unrooted gene trees that were generated by UFboot in Iq-Tree to the analysis of not partitioned loci. Support was assessed as local posterior probability, with ≥ 0.95 considered robust. We also estimated a coalescent-based species tree using *Beast (Heled and Drummond, 2010) in the Beast package (Bouckert *et al.*, 2019), which co-estimate gene trees and species trees from input sequence alignments, and assume that individuals are already correctly assigned to species although the species phylogeny is yet to be estimated (Zhang *et al.*, 2011). We first run the command 'phyluce_align_get_informative_sites' on Phyluce to get the 100 most informative genes, from these we manually chose those that were present in all samples (resulting in 88 UCEs). The analysis was run for 100 million generations sampling every 10.000 generations under a strict clock model with a constant population model, and a Yule model as a tree prior. We used a GTR model (unlinked across loci) for the nucleotide substitution model that was provided by PartitionFinder2 (Lanfear *et al.*, 2017). Tracer v1.7 was used to examine the convergence across runs and the ESS values of sampled parameters. We constructed a maximum clade credibility in TreeAnnotator and visualized the tree using Densitree, both included in the Beast package.

3. RESULTS

3.1. Morphometric geometrics

Regarding the variation in shape of the head, the PCA generated eight Principal Components. The first three components explained 45.28%, 19.16% and 12.37% of the covariance, respectively, totalizing 76.81%. The discriminant analysis correctly assigned 64.4% of the specimens to their respective nominal species and 55.17% to their geographic groups in the cross-validation test. The PCA plot showed superimposition among both groups (Figure 3). Regarding size, the multivariate regression analysis showed that 32.28% of the shape is explained by the allometric effect of the size, with $P < 0.0001$. Nonetheless, even after removing this allometric effect, the groups remained undifferentiated. The comparison between heads centroid sizes showed that there are no significant differences between both nominal species and geographical groups (Figure 3).

The PCA of wings generated 30 Principal Components, of which the first three explained 39.23% of the variation. The first three components explained the 14.59%, 13.89% and 10.75% of the variance, respectively. The cross-validation test correctly located 67% of the nominal species and 61.62% of the geographical groups. A superimposition in PCA plots was also observed in the results of wing's analysis (Figure 4). The regression analysis showed a lower effect of size on shape compared to the ones of the head analysis, 3.59% ($P=0.007$) of variation is explained by the size.

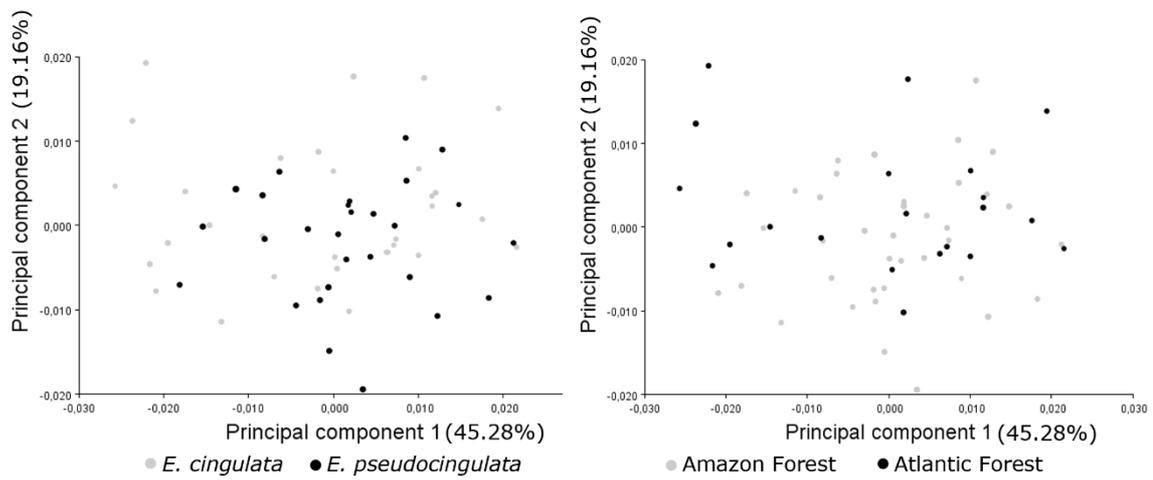


Figure 3. Shape variation of male heads grouped by: **A.** Different nominal species, and **B.** Different geographic groups; by PC 1 and PC 2. The percentage explained by each Principal Components is in parenthesis. The negative and positive extremes of both PC1 and PC2 are shown below and besides of the graph (Factor scale: left -0.03, right 0.03).

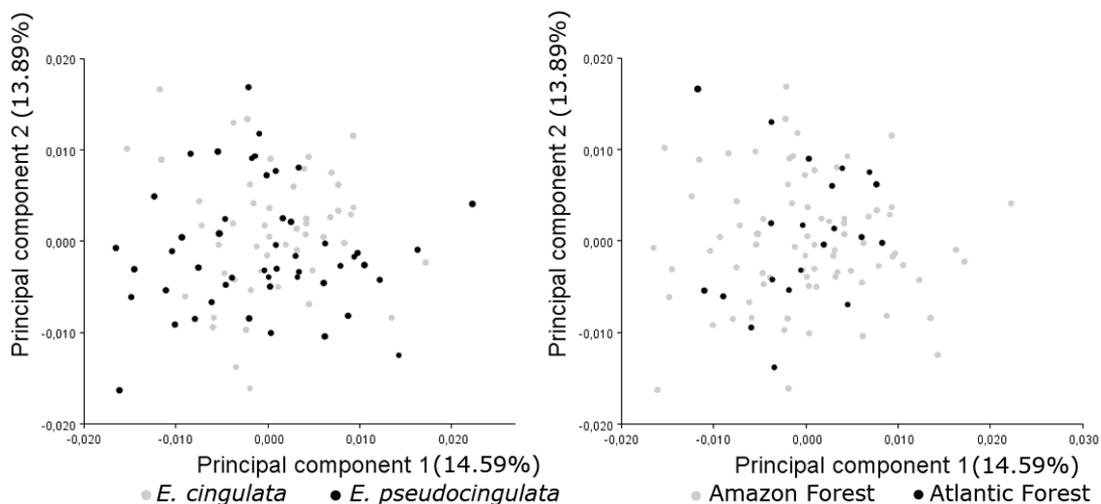


Figure 4. Shape variation of male wings grouped by: **A.** Different nominal species, and **B.**

Different geographic groups; by PC 1 and PC 2. The percentage explained by each Principal Component is in parenthesis. The negative and positive extremes of both PC1 and PC2 are shown below and beside of the graph (Factor scale: left -0.03, right 0.03).

3.2. Genetic distance

The average pairwise genetic p-distance within *E. cingulata* was 1.3% and within *E. pseudocingulata* was 0.7%, while between the two species was 0.95%. This amount of genetic differentiation in mitochondrial DNA indicates no significant difference supporting the presence of two evolutionarily independent units. Additionally, the average pairwise genetic distances among *E. cingulata* and the other analyzed nominal species were between 5.6% and 11.2%, and among *E. pseudocingulata* and the others were between 5.32% and 11.35%. All genetic distances results within and between nominal species are summarized in Table 2.

Table 2. Average pairwise genetic p-distance between nominal species of *Eulaema*.

	<i>Eulaema cingulata</i>	<i>Eulaema pseudocingulata</i>
<i>Eulaema cingulata</i>	1.3%*	0.95%
<i>Eulaema pseudocingulata</i>	0.95%	0.7%*
<i>Eulaema mocsaryi</i>	5.6%	5.32%
<i>Eulaema nigrita</i>	9.3%	9.06%
<i>Eulaema meriana</i>	11.2%	11.35%

*Average pairwise genetic p-distance within each species.

3.3. Phylogenetic relationships

COI sequences were aligned resulting in an alignment of 655 bp. JModel Test identified GTR+I as the best nucleotide substitution model. The consensus tree obtained from MrBayes shows that the COI fragments group *E. cingulata* and *E. pseudocingulata* into one highly supported clade sister of *E. mocsaryi* (Figure 5).

Analysis of the UCE data comprises a total of 21,470,135 reads with an average of 1,431,342 reads per sample (range = 513,594 – 2,585,935). These reads were assembled into an average of 186,700 contigs per sample (range = 34,358 – 441,504), having an average length of 167bp. An average (per sample) of 2,217 of those contigs matched the UCE loci from the target capture probes used. Following the alignment, trimming, and filtering of the UCE contigs, our UCE matrix consisted of 2,180 loci and 1,509,760 bp of sequence data, of which 33,722 bp were informative. The average length of UCE-matching contigs was about 692 bp (range = 229 – 1,831).

The UCEs phylogenetic hypothesis recovered from both partitioned and concatenated schemes used for maximum likelihood inference showed identical topologies, with most nodes receiving high supports (Figures 6.A,B). The group including all specimens of *Eulaema cingulata* and *E. pseudocingulata* was found to be monophyletic with maximum support. The difference between the trees is in the phylogenetic position of *Eulaema pseudocingulata* TA03 and TA04 within a small *pseudocingulata*'s clade. The only individual from the Atlantic Forest (*Eulaema cingulata* TA09) was recovered with maximum support as sister to the individual from Mato Grosso (*Eulaema cingulata* TA07), a state in the Amazon Forest. The ASTRAL and Star Beast species tree had highly support values and both recovered the same topology (Figure 6C).

Even though the COI data included different samples from the UCE data, both methods presented very similar topologies with no significant differences. Both trees showed one clade with the two species included and intermixed with high statistical support. The internal topology resolution show a low divergence levels between these evolutionary units. As we can see in figures 5 and 6, the node corresponding to the species studied showed no separation between nominal species.

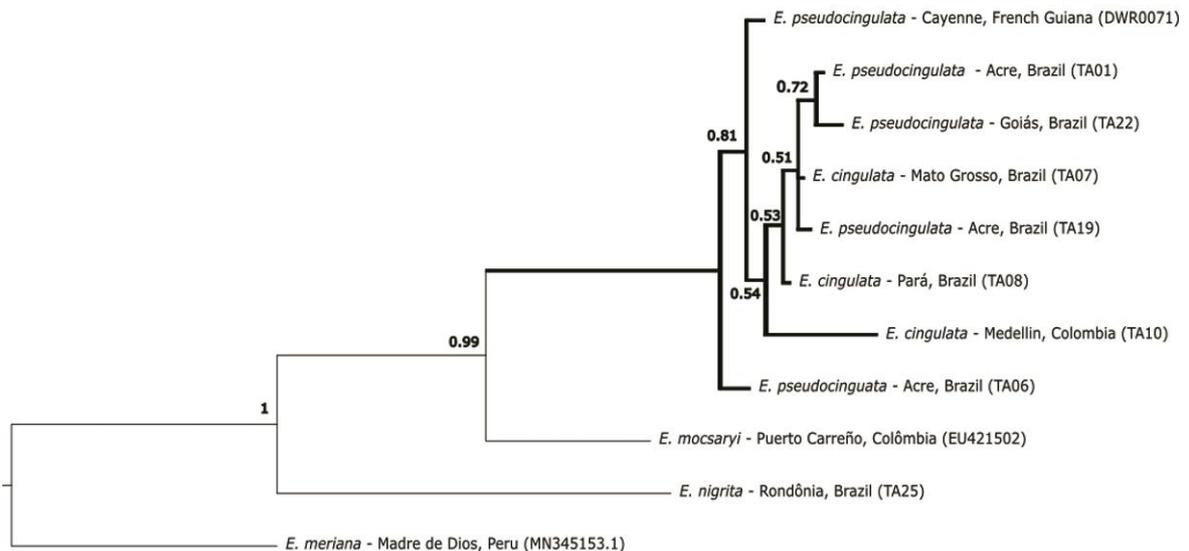


Figure 5: Consensus tree resulting from a Bayesian analysis of molecular data from the gene COI for species of the bee genus *Eulaema*. Posterior probability support is indicated on branches. The codes in front of each analyzed individual correspond to its geographical distribution indicated in Table 1.

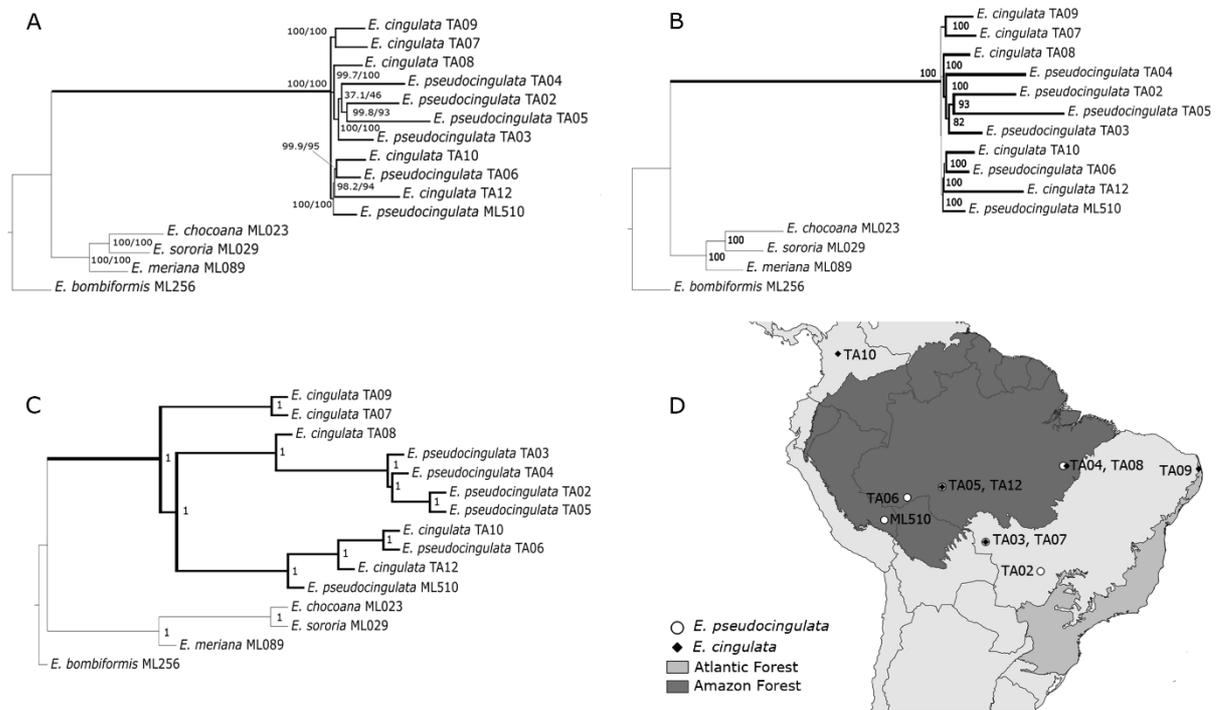


Figure 6. Phylogenetic trees for the nominal *Eulaema cingulata* and *E. pseudocingulata* obtained with UCEs. **A.** Maximum likelihood phylogenetic tree obtained with concatenated dataset of UCEs in Iq-Tree. Numbers on nodes correspond to Ultrafast bootstrap (1000 replicates) and SH-like, respectively. **B.** Maximum likelihood phylogenetic tree obtained with partitioned dataset of UCEs in Iq-Tree. Numbers on nodes correspond to Ultrafast bootstrap (1000 replicates). **C.** Species delimitation analyses based on multispecies-coalescent model obtained with ASTRAL and StarBeast. Number on nodes corresponds to posterior probabilities. **D.** Map showing the distribution of the analyzed material of ingroup.

4. DISCUSSION

A species concept that is in accordance with integrative taxonomy is the General Lineage Concept (GLC), which only required property is that species evolve separately from other lineages (De Queiroz, 2007). Other criteria such as diagnosability, monophyly, or reproductive isolation can be used to recognize separately evolving lineages. The taxonomy of *Eulaema cingulata* and *E. pseudocingulata* has long remained controversial. Based on a multidisciplinary approach, we investigated if these species could be recognized as different taxonomic units. Our morphometric results showed no clustering for the analyzed species neither for different geographic origins, nor among intra or interspecific populations. Geometric

morphometrics of the forewings have been proved as a powerful technique to discriminate bees' species (Francoy *et al.*, 2012; Combey *et al.*, 2013), subspecies (Oleksa & Tofilski, 2014; Silva *et al.*, 2015), cryptic species (Francisco *et al.*, 2008, Hurtado-Burillo *et al.*, 2016) and geographical ecotypes (Francoy *et al.*, 2011; Grassi-Sella *et al.*, 2018, Carneiro *et al.*, 2019). Using both landmarks and outlined-based methodologies, Francoy *et al.* (2012) showed that the use of this approach to discriminate *Euglossa* species was more effective than studies using allozymes and restriction patterns of mitochondrial genes. Quezada-Euán *et al.* (2015) identified differences in wing shape of the two different morphs of *Euglossa viridissima* that had been otherwise identified only by the number of mandibular teeth. The morphometrics study of wings of *Euglossa annectans* Dressler, 1982 and *Euglossa truncata* Rebêlo & Moure, 1995 sampled in the Atlantic Forest and Savannah in Brazil, reported the high morphological similarity between populations albeit the comparatively high dispersal of these bees (Grassi-Sella *et al.*, 2018). We applied the study to the heads in order to assess and compare its power to distinguish among closely related species or forms. The results were congruent with the results obtained with the study of the forewings. Despite being less used than the wings, the use of landmarks on the head has been also informative to recognize intercastes in honeybees (Souza *et al.*, 2015), and to discriminate the morphologically indistinguishable females of the *Psychodopygus* complex (Diptera) (Godoy *et al.*, 2018).

The level of morphological differentiation among recently divergent lineages is sometimes insufficient to recognize species and the use of DNA data has been a useful tool to test the presence of genetically distinct groups of individuals for which one could not find discrete morphological differences. Based on our data, the pair-wise sequence divergences within and between the two species were below 3%, which according to Hebert *et al.* (2003) is a result compatible with the expected variation within a single species. When compared with recognized species, a 'barcode gap' is verified with interspecific distances always larger than 3% (Table 2). Dick *et al.* (2004) found mtDNA divergences within Euglossini species consistently low, with divergences among populations separated by the Andes averaged 1.1%, and 0.4% of mean divergence among cross-Amazon samples (collection sites cover 3.000km). These findings suggest high levels of long-distance gene flow between populations. Rocha-Filho *et al.* (2013) observed a comparatively high dispersal ability in *Eulaema cingulata* through analyses on genetic differentiation between mainland and island populations. López-Uribe *et al.* 2014 also found low values of mitochondrial nucleotide divergence between populations of Euglossini, showing a minimum value of 0.39% in *E. cingulata* among the studied species. According to the authors, the low sequence divergence between populations of

Eulaema cingulata was probably due to the recent origin of this species. The low sequence divergence between individuals of *E. cingulata* found in our study corroborates previous reports, characterizing this bee as displaying large populations with high gene flow.

DNA barcodes are increasingly becoming a standard tool used by taxonomists and its association with morphological characters has proven to help species discrimination in several groups of bees. Although this method has been criticized by some authors (Rubinoff *et al.*, 2006; Wheeler, 2008), it gives additional support to the recognition of species when considered along with other data sources (Padial & De La Riva, 2007). Our phylogenetic analysis yielded an informative tree, providing strong support for the hypothesis that *E. cingulata* and *E. pseudocingulata* belong to the same evolutionary unit. A phylogenetic study using morphological characters recovered these two morphs occupying different species groups within the subgenera *Apeulaema* (Oliveira 2006). The later author found that *E. pseudocingulata* belongs to the group 'peruviana', phylogenetically related to *E. polychroma* while *E. cingulata* belongs to the group 'nigrita' with *E. nigrita*, *E. mocsaryi* and *E. boliviensis*. The first group can be recognized by the lateral projection of the eighth sternum short and rounded, a character that occurs also in the subgenus *Eulaema* (*Eulaema*). This suggest that this character is homoplastic and, consequently, undermines the decision of recognizing species groups within the subgenus *Apeulaema*. The recent revision of the subgeneric classification in *Eulaema* recognized *E. pseudocingulata* as not closely related to *E. polychroma* but as a member of the same species group that *E. cingulata* (Melo, 2014).

The topology obtained with genomic data supports the monophyly and recognition of the clade that comprises *E. cingulata* and *E. pseudocingulata* as previously recognized by COI data. This result corroborate UCEs as another useful tool that provides sufficient variation at shallow time scales to enable species discrimination, adding the utility of these markers for species delimitation studies as observed in other studies (e. g. Smith *et al.*, 2013; Pie *et al.*, 2019; Gueuning *et al.*, 2020). Combined phylogenetic and population genetic approaches were effective to investigate boundaries between complexes of wild European bees suspect to harbor cryptic diversity, mitochondrial introgression, or mitochondrial parphyly (Gueuning *et al.*, 2020). Using COI and UCEs with multispecies coalescent method (BPP), Gueuning *et al.* 2020 noticed that UCEs provided robust species hypotheses and can outperform COI in species delimitation. The adoption of delimitation methods based on the multispecies coalescent model has been criticized by Sukumaran and Knowles (2017), who argued that these methods tend to delimit population structure instead of species. The authors' concern was raised by the possibility of taxonomic inflation if species are described based only on molecular data.

However, most studies using genetic data in species delimitation also incorporate other sources of data and very few studies have been based only on genetic data. The use of species delimitation approaches may increase reliance if different methods offer congruent results within a given dataset.

Herein we conclude that *E. pseudocingulata* does not correspond to an independent evolutionary lineage from *E. cingulata*. Yet, the difference in the shape of the mid tibia velvet area, proposed as the diagnostic character between the morphs, can be indeed interpreted as a variation condition. It can be narrower and farther from the rear edge in some morphs occurring in the Amazon Forest or wider and closer to the rear edge, in morphs occurring throughout the species distribution. Variation is broadly present in bees. Bumble bees, for example, are known for presenting high variability in its coloration pattern (e. g. Carolan *et al.*, 2012; Huang *et al.*, 2015; Koch *et al.*, 2018). Color variation in Eucerini bees was reported by Grando *et al.*, (2018) that found two distinct color patterns in sympatric populations of *Melissodes nigroaenea* in Brazil. Variation in color and shape was observed in *Augochlora amphitrite* by Lepeco & Gonçalves (2018); using morphometric analyses and studying the male genital capsules; the authors did not find any character that could support the recognition of distinct color morphs and macrocephalic females as different species. A study of different color morphs of *Euglossa* species from the Atlantic Forest that were considered as distinct species, received no phylogenetic support, nor enough genetic distance to justify the recognition of the color morphs as different species (Ferrari & Melo, 2014).

We do not know the mechanisms responsible for maintaining this variation in individuals of *Eulaema cingulata* in the Amazon Forest. A similar case was found in a study of two sympatric *Euglossa* species - *E. viridissima* and *E. dilemma*. These are sister species that can be differentiated by the number of mandibular teeth; *E. viridissima* possess two teeth while *E. dilemma* has three with some males of *E. viridissima* expressing a third tooth. These species are distinguished by chemical characters (cuticular hydrocarbons found in the hind tibia) as well as by highly variable DNA markers (microsatellites and SNPs), but show no systematic variation amongst the morphs of *E. viridissima*, confirming that it is in fact a variable species (Eltz *et al.*, 2011; Pokorny *et al.*, 2014; Quezada-Euán *et al.*, 2015). It is difficult to suggest a mechanism for the maintenance of the two morphs in males of *E. cingulata*. However, it is important to understand how genetic diversity is maintained within these populations, and in this sense, further investigations are necessary focusing on the factors that lead to these variations.

5. CONCLUSIONS

We evaluated here possible species limits of *Eulaema cingulata* and *E. pseudocingulata* by integrating multiple independent datasets: geometric morphometrics, mitochondrial DNA, and phylogenomics using ultraconserved elements. All obtained results were congruent regardless of the used methods, showing no separation between morphs that were previously recognized as different species. Our results also suggest that the morphology of the mid tibia of *E. pseudocingulata*, proposed as the diagnostic character between the morphs, is a variable condition of some individuals of *E. cingulata* from the Amazon Forest. Besides the variation in the mid leg, there is also a color variation in some morphs, which causes are yet unknown. Orchid bees are important pollinators in Neotropical forests. In addition to being widely used in environmental quality studies, they have become a good model for evolutionary genetics studies. However, as shown here, there is still the need to ameliorate our knowledge of species delimitation in these bees, even in some of the most ordinary and widespread species in the Neotropical Region.

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